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The Rabbitfish *Siganus virgatus* as Key Macroalgae Browser in Coral Reefs of the Gulf of Thailand

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Abstract: Coral reef resilience is greatly influenced by herbivory. There is a need to identify key fish species fulfilling this critical function in biogeographically distinct regions. This experimental in situ study investigated fish herbivory in coral reefs of the lower Gulf of Thailand characterized by a considerably low herbivorous fish biomass and diversity, but high live coral and low macroalgal cover. This provided an intriguing situation for macroalgal browsing research. Visual census techniques assessed the abundance of local herbivorous fish species, and filmed single-choice assays using the macroalga *Turbinaria* evaluated mass-standardized bites (ms-bites) and biomass removal. Multiple-choice assays offering four locally abundant macroalgae identified specific biomass removal and ms-bites to uncover selection and avoidance patterns of observed fish species. The rabbitfish *Siganus virgatus* constituted only 39% of herbivore biomass but accounted for 90% of ms-bites. In multiple-choice assays, fishes took most (61%) bites on *Sargassum*, followed by *Padina* (28%) and *Turbinaria* (11%), while *Lobophora* was avoided. *S. virgatus* exhibited the most generalized browsing pattern of all species observed. Coinciding with recent studies, our findings suggest that *S. virgatus* plays a key functional role in reefs characterized by low diversity of herbivores and low functional redundancy.

Keywords: Siganus virgatus; macroalgae; browser; herbivorous fish; assay

1. Introduction

Coral reefs constitute the taxonomically most diverse ecosystems on earth [1]. The widespread impact of anthropogenic stressors, compounded by natural disasters, can initiate a cascade of events favoring shifts to macroalgae dominance [1–3]. It is well established that herbivorous fishes play a key role ensuring the continued functioning of reef systems, particularly in the face of negative impacts, due to their ability to remove algae, one of the main competitors of corals [2,4–6]. Nominally herbivorous fishes have traditionally been classified into a small number of distinct functional groups, e.g., based on dietary preferences and methods of food acquisition [7]. Grazers predominantly feed on algal turfs and early life stages of macroalgae, thereby potentially preventing macroalgae establishment [2]. The removal of morphologically and chemically defended mature macroalgae is limited to the functional group of macroalgal browsers [1,7]. Chronic drivers such as overfishing and nutrient pollution can lead to transitional states, with turf-dominated reefs eventually developing into reefs dominated by macroalgae [8,9], but recent literature suggests that turf can also constitute a stable regime in itself [10–12].

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Copyright: © 2021 by the authors. Licensee MDPI, Basel, Switzerland. This article is an open access article distributed under the terms and conditions of the Creative Commons Attribution (CC BY) license (http://creativecommons.org/licenses /by/4.0/). Macroalgae can outgrow feeding pressure when grazing is low, thereby reducing coral settlement, recruitment [13,14] and growth [15], and overgrowing adult corals [16–18]. Similarly, turf algae can inhibit coral larvae settlement and decrease survivorship [19,20]. Increasing reports of coral reefs shifting to macroalgae dominance following large-scale bleaching events [5,21–23] have given rise to a growing number of studies on browsing fishes (e.g., [24–26]; for a full list of Indo-Pacific species see [27]), of which most have been conducted on the Great Barrier Reef (GBR, e.g., [22,25,28]). Key species have been identified based on assessments of estimated fish feeding rates ("estimated function", [29]) or measured loss of macroalgae biomass ("realized function", [29], e.g., [30-32]. However, high variability in the identity of fish species removing macroalgae together with biogeographic differences prevent generalizations [6,33,34], and herbivorous fishes can display nuanced differences in their function across different locations [25,26,30]. Besides physical, chemical, and nutritional traits of the macroalgae coupled with herbivore gut physiology that already exclude most species as potential feeders [24,35–37], numerous factors can influence the feeding selectivity of browsing species, such as temporal and spatial patterns of algal availability [25,28].

On the GBR, studies have repeatedly identified major browsing species to exhibit a strong selection for the brown alga *Sargassum* [26,38,39]. Some of the most consistently important macroalgal browsers in the Indo-Pacific region include *Naso unicornis* [26,39,40], *Kyphosus vaigiensis* [26,36], *Siganus canaliculatus* [38,41] and *Siganus doliatus* [25,42] (see [6]). However, contrasting results exist regarding the two rabbitfish species having an impact functionally equivalent to *N. unicornis* and *K. vaigiensis* in their efficiency of macroalgae removal [38] (but see [43]). Species-specific feeding patterns can lead to varying extents of dietary overlap across a herbivore community [44], resulting in different levels of functional redundancy or complementarity within a reef system [24,30,39]. Pronounced variations in feeding selectivity may be concealed by classifying species into broad functional groups, and assumptions of functional redundancy may be challenged once finer degrees of niche partitioning are considered [42,43,45].

While high diversity systems provide the potential for functional redundancy, the redundancy can be limited in some functional groups [1]. Few species or even a single species can drive macroalgal removal irrespective of the level of herbivore diversity [22,30,39]. Some evidence suggests that species might express greater behavioral flexibility in lower-diversity assemblages, being able to support the function of macroalgal browsing to a similar extent compared to study sites richer in species and thereby increasing resilience [6].

Coral reefs located in the lower Gulf of Thailand such as the ones of the Samui Archipelago provide an interesting research ground for macroalgal browsing, as they exhibit low fish taxonomic diversity and face serious disturbances caused by extensive tourism, overfishing and water pollution [46,47]. To date, little is known about key browsing species including their feeding ecology and functional roles in this region [27]. All study sites exhibited low diversity of herbivorous species and, most strikingly, these reefs are lacking a major component of the herbivorous fish guild–acanthurids [26,39,40]. Based on initial information, herbivorous biomass as well as abundance of invertebrate macroalgae grazers seem to be particularly low as well [48].

The use of remote underwater video cameras has emerged as a viable tool to assess the identity of key species and feeding behavior using transplanted macroalgae in the field [24,26,39], and multiple-choice feeding assays offering different macroalgae species proved a suitable method to assess feeding selectivity of fish species [24,30]. Here, we used visual census techniques to (i) assess the abundance of locally occurring herbivorous fish species, and (ii) filmed macroalgal single-choice assays using the macroalga *Turbinaria* to identify mass-standardized bites (ms-bites) and biomass removal by grazing and browsing fish species. This was supplemented by multiple-choice assays offering locally abundant macroalgae (*Sargassum, Turbinaria, Padina* and *Lobophora*) to (iii) evaluate specific biomass removal and ms-bites to uncover selection and avoidance patterns of observed fish species. The findings of this study contribute to the limited body of information for reefs with low herbivorous fish diversity, particularly in the lower Gulf of Thailand.

2. Materials and Methods

2.1. Study Site

Data were collected at four different sites off the island Ko Pha Ngan, namely Haad Salad (A; 9°47'22.0″ N 99°58'17.1″ E), Mae Haad IV (D; 9°47'32.2″ N 99°58'32.6″ E), Mae Haad II (B; 9°47'47.6″ N 99°58'41.9″ E) and Haad Khom (C; 9°47'53.1″ N 100°00'52.0″ E) (Figure 1), and from a fifth site in the nearby marine park Ang Thong (E; 9°47'55.6″ N 99°42'44.8″ E) located in the lower Gulf of Thailand.

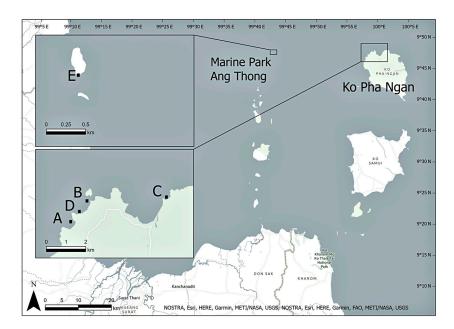


Figure 1. Map of study sites. (A) Haad Salad, (D) Mae Haad IV, (B) Mae Haad II, (C) Haad Khom and (E) Ang Thong Marine Park.

Macroalgae single-choice experiments were conducted between February and June 2015 to assess potential differences on the reef crest of one openly fished site (loc D) and the marine park site (loc E). Macroalgal assays in 2018 were conducted between February and June on the reef flat in three local fished reefs for single-choice assays (loc A–C) and in one fished reef for multiple-choice assays (loc D). The data collection periods covered the dry and part of a transitional season, with southeast monsoons starting from May onwards [49]. The four Ko Pha Ngan study sites (loc A–D) were located on fringing reefs off the north coast of the island subjected to frequent anthropogenic activity [48] and were chosen to investigate macroalgae browsing around the island. At those sites, the reef flat emerges from a sandy beach at a water depth of 1–2 m and stretches into fields of macroalgae composed by the four macroalgal genera *Turbinaria, Sargassum, Padina* and *Lobophora*, coral patches dominated by *Porites sp.* [48], and rubble. Then, it evolves into the reef crest at a water depth of 3–6 m and slopes down to the fore reef at a water depth of 10 m. *Lobophora* was the only macroalga present beyond a water depth of 3 m. The fifth study site (loc E) was situated within a marine protected area, which is a no-take zone by law.

2.2. Fish and Benthic Surveys

For all study sites and years, fish and benthic surveys were carried out every two weeks on the reef crest during the study period in less than 50 m from macroalgae assay deployment sites. The herbivorous fish community was assessed by visual census [7] on SCUBA along three 5 m × 50 m belt transects, yielding a total of 250 m² per transect. Each survey consisted of two divers trained in fish identification swimming left and right from the transect tape recording all fishes identified to species level. Nominally herbivorous fish species were considered in the surveys, including rabbitfishes (Siganidae), parrotfishes (Labridae: Scarinae), surgeonfishes (Acanthuridae), chubs (Kyphosidae), and pomacanthids (genus Pomacanthus) [7]. Individual fish size was estimated and categorized into nine different size classes (0.1–5 cm, 5.1–10 cm, 10.1–15 cm, 15.1–20 cm, 20.1–25 cm, 25.1–30 cm, 30.1–35 cm, 35.1–40 cm and > 40 cm). Herbivorous fish abundance and published length-weight data [50] were used to calculate fish biomass as g m⁻². The benthic community composition was measured using line-point-intercept transects [51] along the same transects used to assess the herbivore community in 0.5 m steps. Benthos was grouped into "live coral cover", "turf algae", "macroalgae", "rubble", "sand" and "other life". Replicates per transect for each surveyed location ranged from 8 to 15 in 2018 and from 9 to 12 in 2015. A period of heavy weather in the beginning of April 2018, with wind speeds of up to 15 knots at the local sites, and difficult accessibility of the marine park resulted in a variation in replicate numbers. A table with replicate numbers of all surveys and macroalgae assays can be found in the supplementary material (Table S1).

2.3. Algae Feeding Assays

In this study, all reported values are stated as mean \pm SE. In 2015, *Turbinaria* appeared to be the most abundant macroalga (Puk, personal observations), which provided the reason for working with this genus in single-choice assays. For consistency, single-choice assays in 2018 also used the then second-most abundant genus *Turbinaria*. As a preparatory assessment, benthic surveys were conducted in February 2018 on the reef flat of the study sites with a replicate number of 3. *Lobophora* was the most abundant macroalga on the reef flat (6 \pm 1%), followed by *Turbinaria* (5 \pm 1%), *Sargassum* (2 \pm 1%) and *Padina* (1 \pm 1%) (see results section). *Sargassum* has been used in various feeding experiments investigating macroalgae browsing (e.g., [26,34]). However, *Sargassum* abundance was very low in both 2015 and 2018 experiments with a seasonal decline until almost completely absent, making the use in assays not feasible [52]. While *Lobophora* was the most abundant alga, its foliose and encrusting morphology makes it difficult to handle in assays [53].

An initial assessment of the diel activity pattern of herbivorous fishes was conducted in February 2018 to determine the most suitable time for macroalgae assays. An underwater video camera of the type GoPro Hero 3+ was deployed after sunrise (7:00) and recorded until sunset (18:30), with battery changes every 2 h. Recordings were analyzed counting the number of herbivorous fishes within each 2 h-block. A Kruskal–Wallis test detected no significant differences in the abundance of fishes between the observed 2 h-blocks (Table S2). Subsequently, assays were performed between 9:00 and 17:00 for 2 h without interruption. Longer recording times would have required battery changes leading to disturbance of potential feeding events, and previous studies using filmed assays have shown the highest bite rate within the first hour of deployment [24,54].

Macroalgae thalli for all assays were collected in adjacent shallow water habitats and kept in a saltwater tank with surface current no longer than 24 h until deployment. Suitable sizes of thalli were chosen, covering the natural variety of different sizes of the respective macroalga found in the study area (see Table S3 for initial weights). Macroalgae thalli were tied onto pieces of coral rubble with fishing line and deployed at the study sites within an area of 1 m². Per assay, one camera was mounted on a tripod built out of PVC and placed in approximately 1 m distance from the used macroalgae to record feeding. To account for biomass loss during 2 h deployment, wet weight of algae was determined to the nearest 0.01 g before and after the assays with an Ohaus Pioneer digital scale. Excess water was removed by spinning algae in a salad spinner for 10 s before weighing. One alga for each genus was covered with a mesh bag and deployed next to each assay, acting as a control accounting for growth and biomass loss not attributed to herbivory (handling

loss). Deployment spots were chosen haphazardly at the respective water depth, with at least 10 m distance to previous spots. A size reference held into the view frame at the beginning of each recording allowed length estimations of feeding fish. Proportionate and absolute biomass loss due to herbivory were calculated accounting for the biomass loss of the control, which simulated biomass loss due to handling: $T_{loss} = 1 - A_a / (A_i \times [1 - H_{loss}])$, where T_{loss} is the proportion of biomass lost, A_a is the algae biomass left after 2 h, A_i is the initial biomass measured before deployment and H_{loss} is the mean percentage of handling loss, calculated based on the controls [25].

Single-choice *Turbinaria* assays conducted in 2015 on the reef crest (3–6 m water depth) were used to compare feeding on Ko Pha Ngan (D; 44 assays and 88 h of video recordings) with the marine park (E; 18 assays and 36 h of video recordings).

In 2018, single-choice assays were placed on the reef flat (1–2 m water depth) in locations A–C (total of 93 assays; 186 h of video recordings) to assess the rate of herbivory in the reef zone in which all four macroalgal genera naturally occurred during the time of the experiment.

Additionally, multiple-choice assays at location D used all four genera of macroalgae (*Turbinaria, Sargassum, Padina* and *Lobophora*) in one setup to investigate selection and avoidance of deployed macroalgae. As *Lobophora* was the only macroalga occurring in water depths beyond 3 m, multiple-choice assays were also conducted on the reef flat (39 assays; 78 h of video recordings). Feeding selectivity was defined by the likelihood that a fish chooses an alga of one genus over another macroalgae genus, if offered on an equal basis [55]. "Proportional-sized" specimens of algae were selected to reflect variation in growth form and size between species at the study site D as required for measures of preference [24] rather than "equal-sized" specimens in terms of biomass and size [54].

2.4. Data Analyses

Video recordings were analyzed for fish species feeding on deployed macroalgae, the size of consumers, and the number of bites taken. Mass-standardized bites (ms-bites) were calculated by multiplying biomass (estimated using fish size and published length-weight relationships, [50]) with observed number of bites (ms-bites = biomass of fish (in kg) × number of bites) following Hoey and Bellwood [39] to account for the herbivory impact of each fish species. Statistical analyses were performed using the software R [56]. Significance level α was 0.05 for all tests. In advance, data were tested for normal distribution using the Kolmogorov-Smirnov test, and homogeneity of variances was tested using Levene's test. Statistical tests were chosen accordingly. Differences in absolute algae biomass before and after feeding assays associated with ms-bites were tested for single-choice assays at site E in 2015 and for multiple-choice assays at location D in 2018 using a Wilcoxon signed-rank test for dependent samples. Remaining locations were not tested due to insufficient sample sizes. A Wilcoxon rank sum test was performed to compare the number of total ms-bites from all fish species together and per fish species separately on single-choice Turbinaria assays at location D with location E in 2015 to test for differences in feeding activity between a local reef and the marine protected area. Correlation between ms-bites from all assays and years and associated proportionate biomass loss was tested using a Spearman correlation test.

Feeding selectivity of individual fish species in multiple-choice assays at location D in 2018 was evaluated according to selection and/or avoidance of offered macroalgae. Positive selection, avoidance, or random selection of offered macroalgae in multiple-choice assays was tested using the Strauss' Linear Selection Index L = ri - pi, where ri is the proportion of the *i*-th food type consumed and pi is the proportion of the *i*-th food type available in the environment [57]. This electivity index was chosen because it is more robust to sampling errors at low r and p values compared to other indices [58]. Due to an overall low number of observed bites, L values could only be calculated for the most dominant browser *S. virgatus*. In addition, the non-parametric Friedman test was performed to test

for a difference in the number of ms-bites taken from the four offered macroalgae *Turbinaria, Sargassum, Padina* and *Lobophora*. The non-parametric Friedman test for dependent samples was chosen because the selection of one macroalgal genus may be dependent on the presence of other algal genera during a trial [59,60]. Further, the Friedman test was used to evaluate the feeding selectivity of the most dominant browser *S. virgatus* according to ms-bites. A Friedman post hoc multiple comparison test according to the Nemenyi method [61] was conducted to perform pairwise comparisons of ms-bites taken by *S. virgatus* on *Turbinaria, Sargassum* and *Padina*.

3. Results

Six rabbitfish species (*Siganus virgatus, Siganus javus, Siganus corallinus, Siganus guttatus, Siganus vermiculatus* and *Siganus stellatus*) and two parrotfish species (*Scarus rivulatus* and *Scarus ghobban*) were observed in all herbivore surveys across years and sites. Sightings of *Naso unicornis, Kyphosus vaigiensis* and *Pomacanthus sexstriatus* occurred rarely (three individuals or less in all assays, ≤2% of biomass).

3.1. Key Browsers and Macroalgae Removal in the Fished vs. Protected Reef

In 2015, the marine park showed a herbivorous fish biomass of 14.1 ± 1.4 g m⁻². *S. rivulatus* made up 36% of herbivorous fish biomass at that site. It was followed by *S. virgatus* (31%), *S. javus* (14%), *S. ghobban* (8%), *S. stellatus* (4%), *S. corallinus* and *S. guttatus* (each 2%) (Figure 2, full details in Table S4). The marine park had a high live coral cover (49 ± 4%), 8 ± 2% of turf algae and no macroalgae (Table 1). The Ko Pha Ngan site D showed less than half of the herbivorous fish biomass of the marine park (6.3 ± 0.6 g m⁻²). Additionally, *S. rivulatus* (59%) and *S. virgatus* (30%) made up most of the herbivore biomass. Benthic community composition showed high live coral cover (43 ± 3%), but also high turf algae cover (32 ± 2%). *Lobophora* was the only macroalga present (12 ± 4%).

Table 1. Benthic community composition on the reef crest in the local reef (loc D) and the marine park (loc E) in 2015. Given are % of total cover for live coral cover (LCC), turf algae, sand, rubble, other life and macroalgae, whereas macroalgae are split into the genera *Turbinara*, *Lobophora*, *Padina* and *Sargassum*. Replicate number n = 12 at loc D and n = 9 at loc E.

2015 Reef Crest		Sir	ngle-Choice		
	Loc	D	Loc E		
Benthos	% Cover	% SE	% Cover	% SE	
LCC	42.75	3.21	48.96	3.85	
Turf Algae	32.44	1.95	7.76	2.09	
Sand	1.34	0.53	20.66	5.08	
Rubble	10.65	3.18	6.34	1.42	
Other Life	0.67	0.3	16.28	2.76	
Macroalgae	12.15	3.51	-	-	
of which					
Turbinaria	-	-	-	-	
Lobophora	12.15	3.51	-	-	
Padina	-	-	-	-	
Sargassum	-	-	-	-	

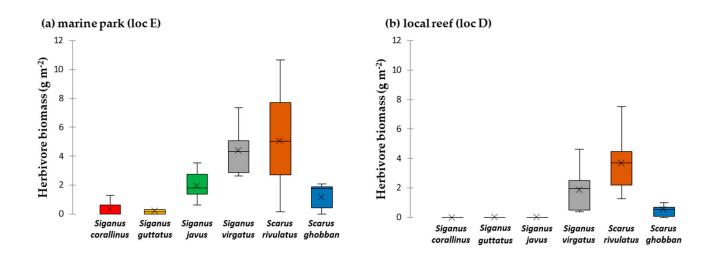


Figure 2. Herbivore biomass surveyed on the reef crest in the marine park (loc E) (**a**) and in a local reef (loc D) (**b**) in 2015. Values are given in mean g m^{-2} . The boxplot shows the median and 25% quantile, the black cross indicates the mean.

In the 2015 single-choice experiments on the reef crest, a total of 438.74 ms-bites during 124 h of recording $(3.54 \pm 0.68 \text{ ms-bites h}^{-1})$ were observed (Figure 3). *S. virgatus* took by far most bites from deployed *Turbinaria* assays, accounting for 93% of all ms-bites (3.29 \pm 0.37 ms-bites h⁻¹), followed by *S. corallinus* (5% of all ms-bites, 0.16 \pm 0.09 ms-bites h⁻¹) and *S. guttatus*, *S. rivulatus*, *S. ghobban* and *S. javus* which took less than 5% of all ms-bites, respectively. The bite rate by all fish species combined was 17-times lower at location D (0.63 \pm 0.36 ms-bites h⁻¹) than at the site E (10.64 \pm 1.14 ms-bites h⁻¹; *p*= 0.002, Table S5). This difference was mainly driven by *S. virgatus*, which took 10.37 \pm 0.71 ms-bites h⁻¹ at location D (*p*< 0.001, Table S5).

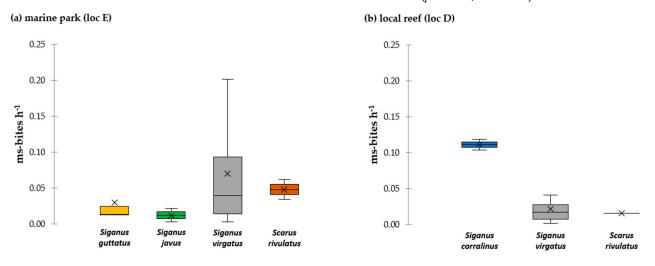


Figure 3. Bite rate (ms-bites h^{-1}) for recorded fish species in single-choice assays on the reef crest in the marine park (loc E) (**a**) and in a local reef (loc D) (**b**) in 2015. Fish species are color-coded. The boxplot shows the median and 25% quantile, the black cross indicates the mean.

3.2. Key Browsers and Macroalgae Removal in Local Reefs

Across the four Ko Pha Ngan sites in 2018, the average herbivorous fish biomass was 10.8 ± 2 g m⁻² and ranged from 3.9 ± 0.2 g m⁻² to 13.8 ± 0.6 g m⁻². *S. virgatus* made up 49% of herbivorous fish biomass. It was followed by *S. rivulatus* (32%), *S. ghobban* (16%), *S. guttatus* (3%), and less than 2% for *S. corallinus*, *S. stellatus* and *S. javus*, respectively (Figure 4, full details in Table S6). Benthos community composition on the reef flat showed high turf algae cover (39 ± 4%), moderate live coral cover (22 ± 7%) and 12 ± 3% macroalgae (Table 2). Benthic data for sites A–D on the reef crest in 2018 was within a similar range as in 2015 for the fished reef (Table S7).

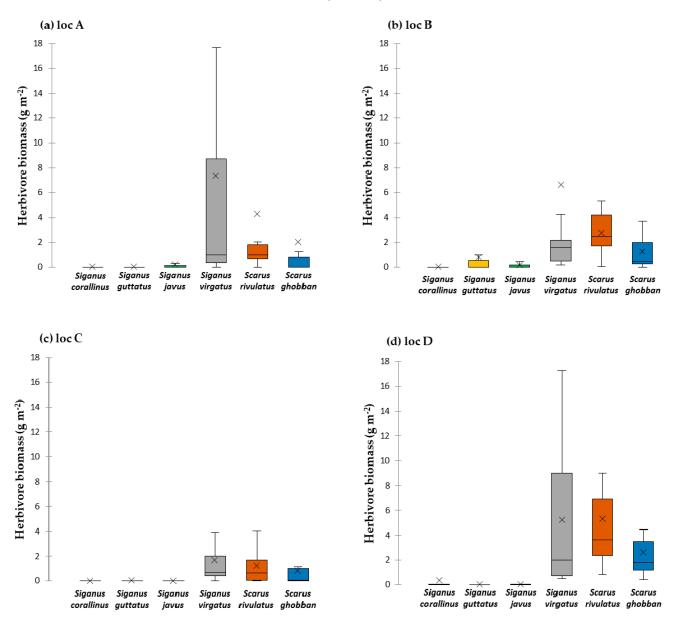


Figure 4. Herbivore biomass surveyed on the reef crest near locations of single-choice assays loc A (**a**), loc B (**b**) and loc C (**c**) and multiple-choice assays loc D (**d**) in 2018. Values are given in mean g m^{-2} . The boxplot shows the median and 25% quantile, the black cross indicates the mean.

Table 2. Benthic community composition surveyed on the reef flat at locations of single-choice (loc A–C) (a) and multiplechoice assays (loc D) (b) in 2018. Given are % of total cover for live coral cover (LCC), turf algae, sand, other life and macroalgae, whereas macroalgae are split into the genera *Turbinara*, *Lobophora*, *Padina* and *Sargassum*. Replicate number n = 3 at each site.

2018 Reef Flat	(a) Single-Choice						(b) Multiple-Choice		
	Loc	Loc A		Loc B		Loc C		Loc D	
Benthos	% Cover	% SE	% Cover	% SE	% Cover	% SE	% Cover	% SE	
LCC	51.8	13.75	7.57	3.27	22.11	7.83	4.95	1.98	
Turf Algae	33.44	6.91	37.5	4.97	56.77	3.81	27.06	1.32	
Sand	0.98	0.57	44.41	3.01	5.61	1.19	54.13	3.68	
Other Life	2.62	1.31	0.66	0.66	0.99	0.57	-	-	
Macroalgae	11.15	3.88	9.87	3.89	14.52	2.35	13.86	2.51	
of which									
Turbinaria	3.61	0.87	4.28	1.32	4.95	1.14	5.61	0.66	
Lobophora	6.23	2.15	4.61	2	8.25	0.33	6.6	1.19	
Padina	1.31	0.87	0.99	0.57	1.32	0.87	-	-	
Sargassum	-	-	-	-	-	-	1.65	0.66	

3.2.1. Single-Choice Assays

While no bites were recorded at locations A and C, a single feeding event of *S. virgatus* was recorded at location B (3.60 ms-bites in 186 h of assays).

3.2.2. Multiple-Choice Assays

Overall, 31.48 ms-bites in 78 h of assays $(0.40 \pm 0.73 \text{ ms-bites h}^{-1})$ were recorded in multiplechoice assays on the reef flat in 2018. The most important macroalgal consumer in all conducted assays was *S. virgatus*, which took 52% of all ms-bites $(0.21 \pm 0.09 \text{ ms-bites h}^{-1})$, followed by *S. guttatus* (19% of all ms-bites, $0.08 \pm 0.32 \text{ ms-bites h}^{-1}$), *S. rivulatus* (12% of all ms-bites, $0.05 \pm 0.17 \text{ ms-bites h}^{-1}$) and *S. corallinus* (11% of all ms-bites, $0.05 \pm 0.14 \text{ ms-bites}$ h⁻¹). *S. ghobban* and *S. javus* took less than 4% of all ms-bites (Figure 5).

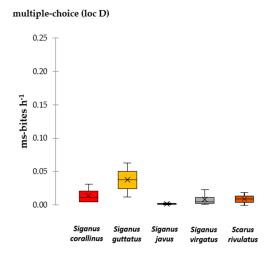


Figure 5. Bite rates (ms-bites h⁻¹) for recorded fish species in multiple choice-assays (loc D) in 2018. Fish species are color-coded. The boxplot shows the median and 25% quantile, the black cross indicates the mean.

3.2.3. Feeding Selection and Avoidance

During the 78 h of recordings, 65% of all ms-bites $(0.26 \pm 0.26 \text{ ms-bites h}^{-1})$ were observed on Sargassum, 25% of all ms-bites were taken on Padina (0.10 ± 0.07 ms-bites h⁻¹), followed by Turbinaria with 10% of all ms-bites $(0.04 \pm 0.48 \text{ ms-bites h}^{-1})$, and Lobophora was avoided in all assays (Figure 6). The Strauss' Linear Selection Index (L) showed a nominal positive selection for *Sargassum* and *Padina*, as well as a slight avoidance of *Turbinaria*, by the most dominant browser S. virgatus (Table 3). The Friedman test revealed a significant selection/avoidance pattern for S. virgatus (p< 0.001, Table S8). Sargassum (55% of all ms-bites) and Padina (41% of all ms-bites) were selected with a similar frequency by S. virgatus, while *Turbinaria* seemed to be less attractive (4% of all ms-bites; *Sargassum - Turbinaria p* = 0.008 and *Padina* - *Turbinaria* p= 0.001, Table S9) and *Lobophora* was always avoided. Due to an overall low number of observed bites, statistical tests could only be performed for the most dominant browser S. virgatus. Based on observations in the field, the siganids were highly selective for Sargassum. S. javus took 86% of all its ms-bites on Sargassum, while S. guttatus and S. corallinus exclusively selected Sargassum. The parrotfish S. rivulatus fed mainly on Turbinaria (73% of all ms-bites exerted by S. rivulatus) but was also observed to take bites on Padina (15% of ms-bites) and Sargassum (12% of ms-bites). S. ghobban was observed to feed on Padina only.

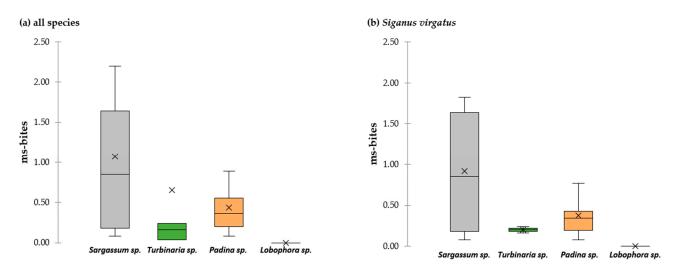


Figure 6. Ms-bites in multiple-choice assays (loc D) in the reef flat in 2018 taken on the four offered macroalgae genera (*Sargassum*, *Padina, Turbinaria* and *Lobophora*) for all fish species (**a**) and for the most dominant browser *S. virgatus* (**b**). No bites on *Lobophora* were recorded. The boxplot shows the median and 25% quantile, the black cross indicates the mean.

Table 3. Strauss' Linear Selection Index (*L*) tested for positive or negative selection of the four offered algae genera *Sargassum, Padina, Turbinaria* and *Lobophora* by *S. virgatus* in multiple-choice assays in the reef flat (loc D) in 2018. *Ri* is the proportion of the *i*-th food type consumed and *Pi* is the proportion of the *i*-th food type available in the environment.

Macroalga	R_i	P_i	L
Sargassum sp.	0.548	0.260	0.003
Padina sp.	0.413	0.105	0.003
Turbinaria sp.	0.039	0.622	-0.006
Lobophora sp.	0.000	0.014	0.000

3.3. Biomass Removal

The difference in biomass for single-choice assays in the marine park (loc E) in 2015 was significant (0.15 ± 0.05 g h⁻¹[$0.2 \pm 0.01\%$ h⁻¹]; p < 0.001, Table S10).

Further, multiple-choice assays conducted on the reef flat (loc D) in 2018 showed a significant difference in absolute biomass loss of 0.04 ± 0.07 g h⁻¹ ($0.38 \pm 0.07\%$ h⁻¹) (p < 0.01, Table S10), whereby *Padina* had the highest removal rate ($0.21 \pm 0.03\%$ h⁻¹), followed by *Sargassum* ($0.15 \pm 0.04\%$ h⁻¹) and *Turbinaria* ($0.03 \pm 0.01\%$ h⁻¹). For *Lobophora*, no biomass loss associated with herbivory was detected. Values for biomass loss are given in Table 4 (2015) and Table 5 (2018).

No correlation was found between ms-bites and measured biomass loss for all assays (Table S11).

Table 4. Biomass loss of macroalgae in single-choice assays in a local reef (loc D) (a) and in the marine park (loc E) (b) on the reef crest in 2015. Given are absolute biomass loss in g $h^{-1} \pm SE$ as well as biomass loss in % $h^{-1} \pm SE$ per algae genus. Locations with significant biomass loss are marked with *.

2015 Reef Crest		Biomass Loss				
	Macroalga	Abs. g h ⁻¹	±SE	% h ⁻¹	±SE	
(a) loc D	Turbinaria sp.	0.066	0.010	0.060	0.001	
(b) loc E *	Turbinaria sp.	0.146	0.053	0.201	0.009	

Table 5. Biomass loss of macroalgae in single-choice assays at loc C (a) and multiple-choice assays at loc D (b) on the reef flat in 2018. Given are absolute biomass loss in g $h^{-1} \pm SE$ as well as biomass loss in % $h^{-1} \pm SE$ per algae genus. Locations with significant biomass loss are marked with *.

2018 Reef Flat		Biomass Loss				
	Macroalga	Abs. g h-1	±SE	% h-1	±SE	
(a) single-choice loc C	Turbinaria sp.	0.003	0.000	-0.018	0.000	
(b) multiple-choice loc D*	Turbinaria sp.	0.013	0.002	0.026	0.007	
	Sargassum sp.	0.009	0.001	0.147	0.035	
	Padina sp.	0.020	0.003	0.209	0.031	
	Lobophora sp.	0.000	0.000	0.000	0.000	

4. Discussion

4.1. Identification of Key Browsers and Macroalgae Removal

The herbivorous fish biomass at all sites of this study was lower than the average herbivore biomass in locations accessible to fisheries throughout the Indo-Pacific (10.8 ± 2 g m⁻² vs. 29.0 g m⁻², [62,63]). Fishing has been observed at all sites including the marine park as the fishing ban is typically not enforced. The group of rabbitfishes made up 53% of total herbivore biomass on average, and *S. virgatus* alone constituted 49% of that biomass. Parrotfishes comprised 47% of the herbivore community, whereas surgeonfishes, which constitute an important group of herbivores in other reef systems [33,39,64], were absent. *S. virgatus* was the single most important browser observed at all study locations, exerting 90% of all ms-bites at all locations combined, ranging from 97% at location E and 62% at location D in single-choice assays in 2015 to 52% in multiple-choice assays in 2018.

The single-choice experiments aimed to gain insight into key herbivores and to assess rates of herbivory. The bite rate on the reef crest of location D of Ko Pha Ngan was 17 times lower than at the marine park site and was driven largely by *S. virgatus*. Additionally, higher background densities of algae at the Ko Pha Ngan sites may have deterred feeding [65,66]. In turn, the deployed *Turbinaria* at the protected marine park site E, which harbored turf algae in low abundance (<10%) and no macroalgae, appeared to have been an attractive subsidy. Feeding on the *Turbinaria* algae from the assays may be beneficial compared to the day-to-day foraging strategy of herbivorous fish, as the effort needed to crop small algae in complex reef habitats entails a significant risk of predation and/or a lower feeding efficiency [67,68]. Transplanted assays are also likely to be more apparent

and may attract browsers from beyond their normal foraging range [64]. Biomass removal at location E was significant ($0.2 \pm 0.01\%$ h⁻¹).

The single-choice experiments on the reef flat, the reef zone with the highest macroalgae cover and in which all genera were naturally occurring, showed virtually no feeding. This is in line with previous studies, in which significantly lower rates of herbivory compared to other reef habitats were recorded on the reef flat (e.g., [69]). The reef crest generally has the highest diversity of fishes [70,71], extensive territoriality with fishes protecting preferred feeding locations [72], the highest rates of primary productivity [73,74], and it often is the preferred feeding habitat for macroalgae browsers [42,69]. The overall high availability of algae in the Ko Pha Ngan reef flats likely led to a low attractiveness of the algae assays compared to the surrounding, resulting in diluted feeding pressure as assays are likely to be less apparent [25,64,75]. Similarly, macroalgal removal may decrease with higher macroalgal density due to higher predation risk and/or the reduced ability of herbivores to detect predators [65,69,76]. However, variation in initial algae biomass in single-choice assays may have influenced feeding rates as resource quantity can influence herbivory [77], hence leading to a difference in observed patterns between the singlechoice assays at different sites.

Turbinaria was selected for the single-choice assays due to its higher relative abundance compared to other macroalgae and because of the low seasonal abundance of *Sargassum*, which is regularly used in feeding assays. However, the use of *Turbinaria* could be an explanation for the rather low feeding rates in our study, since it is usually not preferred by herbivores such as *S. virgatus* as it has a rather low palatability due to several morphological and chemical defenses [78–80] and is mainly consumed by large herbivores such as *N. unicornis* and *K. vaigiensis* not present in this study [30,36,42]. We therefore suggest further feeding assessments using single-choice assays of *Sargassum* as well as of *Turbinaria* during different times of the year to uncover possible variations in herbivory linked to macroalgae genus and algae condition in the studied reefs [52,81,82].

The capability to digest macroalgae varies between herbivores and may determine the benefit gained by feeding on them. The analysis of short chain fatty acid (SCFA) levels of different herbivores places siganids in between browsers, such as kyphosids and acanthurids feeding on large macroscopic algae including fucoids, and scarine parrotfishes which feed on detritus, indicating a mixed diet with algal components [83]. The three siganid species observed at the study sites, *S. guttatus, S. corallinus* and *S. javus*, may not have been capitalizing on the subsidies due to different feeding preferences, as they are considered omnivorous [41,84,85] or primarily feed on turf [86]. All assays tested the consumption of macroalgae only. Since turf algae is a major component of coral reef systems, further investigation to elucidate grazing on turf algae on the flat and crest of the studied reefs would yield interesting results.

The browsing rabbitfish *S. virgatus* contributed 97% of mass-standardized bites at site E. *S. virgatus* is a mixed herbivore feeding on both macroalgae and the epilithic algal matrix [87]. In recent years several studies have emerged on the role of *S. virgatus* contributing towards macroalgal (i.e., *Sargassum*) removal within various ecosystems in the central Indo-Pacific region, and it has been concluded that *S. virgatus* is the most functionally important macroalgal remover along a gradient of reef degradation, under varying environmental conditions and over different spatial and temporal scales [34,76,82,88].

Our results clearly show that *S. virgatus* thrives in surroundings with virtually no edible macroalgae present, such as the reef flat of this study, feeding on turf algae and small thalli of macroalgae species, but is also capable of consuming adult macroalgae stands. *S. virgatus* is closely related to *S. doliatus*, and their division into separate species is considered incipient [89]. Although *S. doliatus* has been previously identified as an algal cropper [41,70,90] and its functional impact has been questioned [41,43], *S. doliatus* is capable of consuming adult macroalgal stands and its abundance in many inshore regions of the GBR has led to the suggestion that macroalgal browsing in those reefs might be reliant on *S. doliatus* as a single species [91].Like its sister species *S. doliatus*, *S. virgatus* is a

mixed herbivore with high functional plasticity, and its prevalence and feeding habits may prove useful in low-diversity systems to sustain important functions in the absence of more specialized herbivores such as *K. vaigiensis* and *N. unicornis* [41,42].

4.2. Selection and Avoidance of Macroalgae in Multiple-Choice Assays

S. virgatus and *S. rivulatus* displayed a broader feeding pattern in multiple-choice assays at location D, consuming three out of four offered macroalgae, while the other herbivores observed were highly selective. Biomass loss in multiple-choice assays $(0.4 \pm 0.01\%$ h⁻¹) was considerably low [54]. Only a few studies have been able to find a relationship between feeding intensity and biomass loss of macroalgae ([39], but see [38]). However, the variation of initial weights in algae assays needs to be considered as well, as it can influence the rate of herbivory and lead to lower biomass removal [77].

In line with previous observations [24,34,41], the brown alga Sargassum was most readily consumed in this study (65% of all ms-bites and 0.15 ± 0.04% h⁻¹ biomass loss) and selected for by five out of six species, including all siganids. S. virgatus fed on Sargassum (55% of ms-bites), Padina (41% of ms-bites) and to a smaller extent on Turbinaria (4% of ms-bites), while the remaining siganid species almost exclusively selected Sargassum. Only 10% of all ms-bites in multiple-choice assays and a biomass loss of $0.03 \pm 0.01\%$ h⁻¹ were observed on Turbinaria, which was targeted mainly by S. rivulatus (73% of ms-bites). Parrotfishes such as S. rivulatus and S. ghobban can readily consume macroalgae with strong structural defenses [36,92], and they consume various substrates to obtain their main food source endo- and epilithic microbial resources. Despite few reports of parrotfish species feeding on macroalgae on the GBR [24,54], macroalgae removal is considered incidental [37,83,93]. Epibionts living on Turbinaria constitute a secondary food source for browsing species [94], and epiphyte loads are known to shape feeding patterns [28]. Since epibionts were observed in the assays, there is a high likelihood that both S. rivulatus and S. virgatus were targeting epibionts in our assays, rather than Turbinaria itself, leading to incidental low biomass removal and explaining why no relationship between ms-bites and biomass loss could be found [22,28,38,90].

S. ghobban fed on *Padina* only, which was the second most targeted alga (25% of all ms-bites) and showed the highest biomass loss ($0.2 \pm 0.03\%$ h⁻¹). *Lobophora* was highly abundant at the study sites and no herbivory was observed in our assays. Indeed, most herbivore species are not able to derive nutrients from *Lobophora* [36,83,95] and a recent study showed different grazing fishes removing *Lobophora* recruits to varying degrees, while removal of adult *Lobophora* was very low [96].

Our study found low removal of algal biomass by herbivores on the reef crest and almost none on the reef flat, which questions whether the essential ecosystem function of macroalgal removal (the "realized function" [29]) in the fished reefs is maintained. However, the single-choice assays investigated feeding on Turbinaria, which is strongly defended and usually not a preferred macroalga [80]. As indicated by the results of our multiple-choice assays, at least two of the other occurring macroalgae genera are likely to be effectively removed from the reef crest, as corroborated by the observation that only Lobophora occurred at depths >3 m. The local reefs in our study showed a relatively high live coral cover and low macroalgae cover, indicating that herbivores successfully prevent macroalgae establishment by grazing on algal turfs and early life stages of macroalgae. However, further investigation is crucially needed to quantify overall grazing pressure of the fish species across different habitats. Our results corroborate the role of *S. virgatus* as an opportunistic herbivore, thriving in environments with varying turf algae and macroalgae availability and feeding on a range of algae, and hence demonstrating functional plasticity in response to environmental and biological variations. S. virgatus displayed a generalized feeding pattern in a herbivore community characterized by low biomass and diversity and consisting mostly of specialized species. S. virgatus appears to constitute a key consumer species in the study region and likely is of great importance in herbivore-depauperate reef systems such as the investigated sites.

Previous studies found the functional role of species to be highly context dependent [24], e.g., differing between locations and depending on co-occurring species [34,86]. Potential functional groups may be revealed in experiments, but anthropogenic and natural disturbances increasingly alter coral reef systems towards macroalgae being a dominant component [93,97]. The potential of herbivore species to exert critical functional roles in benthic communities, such as macroalgae browsing, may only be fully exposed when conditions change, and might be essential for coral reef resilience in the present and future.

5. Conclusions

Currently, the ecosystem function of macroalgae control seems sufficiently supported by the herbivorous fish community in the studied reefs. Based on the important role *S. virgatus* appears to play on the studied reefs, it should be given particular consideration in spatial and species-specific management efforts. This could be relevant to ensure provision of ecosystem functions and to bolster the resilience of reefs harboring species differing from the most consistently identified browsing species in times of frequent overfishing and increased anthropogenic disturbances.

Supplementary Materials: The following are available online at www.mdpi.com/1424-2818/13/3/123/s1, Table S1: study design, Table S2: diel assays, Table S3: Initial algae weights, Table S4: fish census data 2015, Table S5: Wilcoxon Test for ms-bites, Table S6: fish census data 2018, Table S7: benthos reef crest 2018, Table S8: Friedman Test, Table S9: Friedman post hoc Nemenyi Test, S10: Wilcoxon Test for biomass loss, S11: Spearman correlation, S12 – ms-bites and raw bites in multiple-and single-choice assays, S13-S16: raw data.

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References

1. Bellwood, D.R.; Hughes, T.P.; Folke, C.; Nyström, M. Confronting the coral reef crisis. *Nat. Cell Biol.* 2004, 429, 827–833, doi:10.1038/nature02691.

- Hughes, T.P.; Rodrigues, M.J.; Bellwood, D.R.; Ceccarelli, D.; Hoegh-Guldberg, O.; McCook, L.; Moltschaniwskyj, N.; Pratchett, M.S.; Steneck, R.S.; Willis, B. Phase shifts, herbivory, and the resilience of coral reefs to climate change. *Curr. Biol.* 2007, 17, 360– 365, doi:10.1016/j.cub.2006.12.049.
- Mumby, P.J. Phase shifts and the stability of macroalgal communities on Caribbean coral reefs. Coral Reefs 2009, 28, 761–773, doi:10.1007/s00338-009-0506-8.
- 4. Mumby, P.J.; Harborne, A.R. Marine reserves enhance the recovery of corals on Caribbean reefs. *PLoS ONE* **2010**, *5*, e8657, doi:10.1371/journal.pone.0008657.
- 5. Graham, N.A.J.; Jennings, S.; MacNeil, M.A.; Mouillot, D.; Wilson, S.K. Predicting climate-driven regime shifts versus rebound potential in coral reefs. *Nat. Cell Biol.* **2015**, *518*, 94–97, doi:10.1038/nature14140.
- 6. Tebbett, S.B.; Hoey, A.S.; Depczynski, M.; Wismer, S.; Bellwood, D.R. Macroalgae removal on coral reefs: Realised ecosystem functions transcend biogeographic locations. *Coral Reefs* **2019**, *39*, 203–214, doi:10.1007/s00338-019-01874-w.
- Monitoring functional groups of herbivorous reef fishes as indicators of coral reef resilience. In A Practical Guide for Coral Reef Managers in the Asia Pacific Region; Green, A.L., Bellwood, D.R., Eds.; 2009 ed.; International Union for Conservation of Nature: Gland, Switzerland, 2009. ISBN 283171169X.
- 8. Done, T.J. Phase shifts in coral reef communities and their ecological significance. *Hydrobiologia* **1992**, 247, 121–132, doi:10.1007/bf00008211.
- 9. Hughes, T.P. Catastrophes, phase shifts, and large-scale degradation of a caribbean coral reef. *Science* **1994**, *265*, 1547–1551, doi:10.1126/science.265.5178.1547.
- Jouffray, J.-B.; Nyström, M.; Norström, A.V.; Williams, I.D.; Wedding, L.M.; Kittinger, J.N.; Williams, G.J. Identifying multiple coral reef regimes and their drivers across the Hawaiian archipelago. *Philos. Trans. R. Soc. B Biol. Sci.* 2015, 370, 20130268, doi:10.1098/rstb.2013.0268.
- Donovan, M.K.; Friedlander, A.M.; Lecky, J.; Jouffray, J.-B.; Williams, G.J.; Wedding, L.M.; Crowder, L.B.; Erickson, A.L.; Graham, N.A.J.; Gove, J.M.; et al. Combining fish and benthic communities into multiple regimes reveals complex reef dynamics. *Sci. Rep.* 2018, *8*, 16943, doi:10.1038/s41598-018-35057-4.
- Jouffray, J.-B.; Wedding, L.M.; Norström, A.V.; Donovan, M.K.; Williams, G.J.; Crowder, L.B.; Erickson, A.L.; Friedlander, A.M.; Graham, N.A.J.; Gove, J.M.; et al. Parsing human and biophysical drivers of coral reef regimes. *Proc. R. Soc. B Boil. Sci.* 2019, 286, 20182544, doi:10.1098/rspb.2018.2544.
- 13. Kuffner, I.B.; Walters, L.J.; Becerro, M.A.; Paul, V.J.; Ritson-Williams, R.; Beach, K.S. Inhibition of coral recruitment by macroalgae and cyanobacteria. *Mar. Ecol. Prog. Ser.* 2006, *323*, 107–117, doi:10.3354/meps323107.
- 14. Mumby, P.J.; Steneck, R.S.; Adjeroud, M.; Arnold, S.N. High resilience masks underlying sensitivity to algal phase shifts of Pacific coral reefs. *Oikos* **2016**, *125*, 644–655, doi:10.1111/oik.02673.
- 15. Box, S.; Mumby, P. Effect of macroalgal competition on growth and survival of juvenile Caribbean corals. *Mar. Ecol. Prog. Ser.* **2007**, *342*, 139–149, doi:10.3354/meps342139.
- 16. Lirman, D. Competition between macroalgae and corals: Effects of herbivore exclusion and increased algal biomass on coral survivorship and growth. *Coral Reefs* **2001**, *19*, 392–399, doi:10.1007/s003380000125.
- 17. Jompa, J.; McCook, L.J. Effects of competition and herbivory on interactions between a hard coral and a brown alga. *J. Exp. Mar. Biol. Ecol.* **2002**, *271*, 25–39, doi:10.1016/s0022-0981(02)00040-0.
- 18. Vieira, C.; Payri, C.; De Clerck, O. Overgrowth and killing of corals by the brown alga L obophora hederacea (Dictyotales, Phaeophyceae) on healthy reefs in New Caledonia: A new case of the epizoism syndrome. *Psychol. Res.* **2015**, *63*, 152–153.
- Birrell, C.L.; McCook, L.J.; Willis, B.L. Effects of algal turfs and sediment on coral settlement. Mar. Pollut. Bull. 2005, 51, 408–414, doi:10.1016/j.marpolbul.2004.10.022.
- Arnold, S.N.; Steneck, R.S.; Mumby, P.J. Running the gauntlet: Inhibitory effects of algal turfs on the processes of coral recruitment. *Mar. Ecol. Prog. Ser.* 2010, 414, 91–105, doi:10.3354/meps08724.
- Hughes, T.P.; Baird, A.H.; Bellwood, D.R.; Card, M.; Connolly, S.R.; Folke, C.; Grosberg, R.; Hoegh-Guldberg, O.; Jackson, J.B.C.; Kleypas, J.; et al. Climate Change, Human Impacts, and the Resilience of coral reefs. *Science* 2003, 301, 929–933, doi:10.1126/science.1085046.
- 22. Bellwood, D.R.; Hughes, T.P.; Hoey, A.S. Sleeping functional group drives coral-reef recovery. *Curr. Biol.* 2006, *16*, 2434–2439, doi:10.1016/j.cub.2006.10.030.
- 23. Hughes, T.P.; Kerry, J.T.; Simpson, T. Large-scale bleaching of corals on the Great Barrier Reef. *Ecology* 2018, 99, 501, doi:10.1002/ecy.2092.
- 24. Mantyka, C.S.; Bellwood, D.R. Macroalgal grazing selectivity among herbivorous coral reef fishes. *Mar. Ecol. Prog. Ser.* 2007, 352, 177–185, doi:10.3354/meps07055.
- 25. Bennett, S.; Bellwood, D. Latitudinal variation in macroalgal consumption by fishes on the Great Barrier Reef. *Mar. Ecol. Prog. Ser.* **2011**, *426*, 241–252, doi:10.3354/meps09016.
- Vergés, A.; Bennett, S.; Bellwood, D.R. Diversity among macroalgae-consuming fishes on coral reefs: A transcontinental comparison. *PLoS ONE* 2012, 7, e45543, doi:10.1371/journal.pone.0045543.
- 27. Puk, L.D.; Ferse, S.C.A.; Wild, C. Patterns and trends in coral reef macroalgae browsing: A review of browsing herbivorous fishes of the Indo-Pacific. *Rev. Fish Biol. Fish.* **2015**, *26*, 53–70, doi:10.1007/s11160-015-9412-z.
- 28. Lefèvre, C.D.; Bellwood, D.R. Temporal variation in coral reef ecosystem processes: Herbivory of macroalgae by fishes. *Mar. Ecol. Prog. Ser.* **2011**, 422, 239–251, doi:10.3354/meps08916.

- 29. Bellwood, D.R.; Streit, R.P.; Brandl, S.J.; Tebbett, S.B. The meaning of the term 'function'in ecology: A coral reef perspective. *Funct. Ecol.* **2019**, *33*, 948–961.
- Rasher, D.B.; Hoey, A.S.; Hay, M.E. Consumer diversity interacts with prey defenses to drive ecosystem function. *Ecology* 2013, 94, 1347–1358, doi:10.1890/12-0389.1.
- Bonaldo, R.M.; Pires, M.M.; Guimarães, P.R.; Hoey, A.S.; Hay, M.E. Small marine protected areas in Fiji provide refuge for reef fish assemblages, feeding groups, and corals. *PLoS ONE* 2017, *12*, e0170638, doi:10.1371/journal.pone.0170638.
- Loffler, Z.; Hoey, A.S. Canopy-forming macroalgal beds (Sargassum) on coral reefs are resilient to physical disturbance. J. Ecol. 2018, 106, 1156–1164, doi:10.1111/1365-2745.12875.
- Cvitanovic, C.; Bellwood, D.R. Local variation in herbivore feeding activity on an inshore reef of the Great Barrier Reef. Coral Reefs 2008, 28, 127–133, doi:10.1007/s00338-008-0433-0.
- Plass-Johnson, J.G.; Ferse, S.C.A.; Jompa, J.; Wild, C.; Teichberg, M. Fish herbivory as key ecological function in a heavily degraded coral reef system. *Limnol. Oceanogr.* 2015, 60, 1382–1391, doi:10.1002/lno.10105.
- 35. Paul, V.; Hay, M. Seaweed susceptibility to herbivory: Chemical and morphological correlates. *Mar. Ecol. Prog. Ser.* **1986**, *33*, 255–264, doi:10.3354/meps033255.
- 36. Choat, J.H.; Robbins, W.D.; Clements, K.D. The trophic status of herbivorous fishes on coral reefs. *Mar. Biol.* 2004, 145, 445–454, doi:10.1007/s00227-004-1341-7.
- Clements, K.D.; Raubenheimer, D.; Choat, J.H. Nutritional ecology of marine herbivorous fishes: Ten years on. *Funct. Ecol.* 2009, 23, 79–92, doi:10.1111/j.1365-2435.2008.01524.x.
- 38. Fox, R.J.; Bellwood, D.R. Remote video bioassays reveal the potential feeding impact of the rabbitfish Siganus canaliculatus (f: Siganidae) on an inner-shelf reef of the Great Barrier Reef. *Coral Reefs* **2008**, *27*, 605–615, doi:10.1007/s00338-008-0359-6.
- 39. Hoey, A.S.; Bellwood, D.R. Limited functional redundancy in a high diversity system: Single species dominates key ecological process on coral reefs. *Ecosystems* **2009**, *12*, 1316–1328, doi:10.1007/s10021-009-9291-z.
- 40. Ford, A.K.; Bejarano, S.; Marshell, A.; Mumby, P.J. Linking the biology and ecology of key herbivorous unicornfish to fisheries management in the Pacific. *Aquat. Conserv. Mar. Freshw. Ecosyst.* **2016**, *26*, 790–805, doi:10.1002/aqc.2623.
- 41. Hoey, A.S.; Brandl, S.J.; Bellwood, D.R. Diet and cross-shelf distribution of rabbitfishes (f. Siganidae) on the northern Great Barrier Reef: Implications for ecosystem function. *Coral Reefs* **2013**, *32*, 973–984, doi:10.1007/s00338-013-1043-z.
- Loffler, Z.; Bellwood, D.R.; Hoey, A.S. Among-habitat algal selectivity by browsing herbivores on an inshore coral reef. Coral Reefs 2015, 34, 597–605, doi:10.1007/s00338-015-1265-3.
- 43. Streit, R.P.; Hoey, A.S.; Bellwood, D.R. Feeding characteristics reveal functional distinctions among browsing herbivorous fishes on coral reefs. *Coral Reefs* **2015**, *34*, 1037–1047, doi:10.1007/s00338-015-1322-y.
- Kelly, E.L.A.; Eynaud, Y.; Clements, S.M.; Gleason, M.; Sparks, R.T.; Williams, I.D.; Smith, J.E. Investigating functional redundancy versus complementarity in Hawaiian herbivorous coral reef fishes. *Oecologia* 2016, *182*, 1151–1163, doi:10.1007/s00442-016-3724-0.
- 45. Brandl, S.J.; Bellwood, D.R. Individual-based analyses reveal limited functional overlap in a coral reef fish community. *J. Anim. Ecol.* **2014**, *83*, 661–670, doi:10.1111/1365-2656.12171.
- 46. Cheevaporn, V.; Menasveta, P. Water pollution and habitat degradation in the Gulf of Thailand. *Mar. Pollut. Bull.* **2003**, 47, 43–51, doi:10.1016/s0025-326x(03)00101-2.
- 47. Adger, W.N.; Hughes, T.P.; Folke, C.; Carpenter, S.R.; Rockström, J. Social-ecological resilience to coastal disasters. *Science* 2005, 309, 1036–1039, doi:10.1126/science.1112122.
- Stuhldreier, I.; Bastian, P.; Schönig, E.; Wild, C.; Schoenig, E. Effects of simulated eutrophication and overfishing on algae and invertebrate settlement in a coral reef of Koh Phangan, Gulf of Thailand. *Mar. Pollut. Bull.* 2015, 92, 35–44, doi:10.1016/j.marpolbul.2015.01.007.
- 49. Buranapratheprat, A.; Bunpapong, M. A two dimensional hydrodynamic model for the Gulf of Thailand. In Proceedings of the IOC/WESTPAC Fourth International Scientific Symposium, Okinawa, Japan, 2 February 1998; Volume 469, p. 478.
- Kulbicki, M.; Guillemot, N.; Amand, M. A general approach to length-weight relationships for New Caledonian lagoon fishes. *Cybium* 2005, 29, 235–252.
- Hill, J.; Wilkinson, C.R. Methods for ecological monitoring of coral reefs. In A Resource for Managers, Version 1; Australian Institute of Marine Science: Townsville, Australia, 2004. ISBN 9780642322371.
- 52. Low, J.K.Y.; Fong, J.; Todd, P.A.; Chou, L.M.; Bauman, A.G. Seasonal variation of *Sargassum ilicifolium* (phaeophyceae) growth on equatorial coral reefs. *J. Phycol.* **2018**, *55*, 289–296, doi:10.1111/jpy.12818.
- Vieira, C.; D'Hondt, S.; De Clerck, O.; Payri, C.E. Toward an inordinate fondness for stars, beetles and Lobophora? Species diversity of the genus Lobophora (Dictyotales, Phaeophyceae) in New Caledonia. J. Phycol. 2014, 50, 1101–1119, doi:10.1111/jpy.12243.
- 54. Mantyka, C.S.; Bellwood, D.R. Direct evaluation of macroalgal removal by herbivorous coral reef fishes. *Coral Reefs* **2007**, *26*, 435–442, doi:10.1007/s00338-007-0214-1.
- 55. Davis, A.; Manly, B.; McDonald, L.; Thomas, D. Resource selection by animals: Statistical design and analysis for field studies. *J. Anim. Ecol.* **1994**, *63*, 745, doi:10.2307/5247.
- 56. R Core Team. R: A Language and Environment for Statistical Computing; R Foundation for Statistical Computing: Vienna, Austria, 2018. Available online: https://www.R-project.org (accessed on 26 06 2020).

- 57. Strauss, R.E. Reliability estimates for Ivlev's electivity index, the forage ratio, and a proposed linear index of food selection. *Trans. Am. Fisheries Soc.* **1979**, *108*, 344–352.
- 58. Lechowicz, M.J. The sampling characteristics of electivity indices. *Oecologia* **1982**, *52*, 22–30, doi:10.1007/bf00349007.
- 59. Roa, R. Design and analysis of multiple-choice feeding-preference experiments. *Oecologia* **1992**, *89*, 509–515, doi:10.1007/bf00317157.
- Lockwood, J.R., III On the statistical analysis of multiple-choice feeding preference experiments. *Oecologia* 1998, 116, 475–481, doi:10.1007/s004420050612.
- Hollander, M.; Wolfe, D.A.; Chicken, E. Nonparametric Statistical Methods, 3rd ed.; John Wiley & Sons, Inc.: Hoboken, NJ, USA, 2014. ISBN 0470387378.
- 62. Roff, G.; Mumby, P.J. Global disparity in the resilience of coral reefs. *Trends Ecol. Evol.* 2012, 27, 404–413, doi:10.1016/j.tree.2012.04.007.
- Edwards, C.B.; Friedlander, A.M.; Green, A.G.; Hardt, M.J.; Sala, E.; Sweatman, H.P.; Williams, I.D.; Zgliczynski, B.; Sandin, S.A.; Smith, J.E. Global assessment of the status of coral reef herbivorous fishes: Evidence for fishing effects. *Proc. R. Soc. B Boil. Sci.* 2014, 281, 20131835, doi:10.1098/rspb.2013.1835.
- 64. Chong-Seng, K.M.; Nash, K.L.; Bellwood, D.R.; Graham, N.A.J. Macroalgal herbivory on recovering versus degrading coral reefs. *Coral Reefs* **2014**, *33*, 409–419, doi:10.1007/s00338-014-1134-5.
- Hoey, A.S.; Bellwood, D.R. Suppression of herbivory by macroalgal density: A critical feedback on coral reefs? *Ecol. Lett.* 2011, 14, 267–273, doi:10.1111/j.1461-0248.2010.01581.x.
- Madin, E.M.P.; Gaines, S.D.; Warner, R.R. Field evidence for pervasive indirect effects of fishing on prey foraging behavior. *Ecology* 2010, *91*, 3563–3571, doi:10.1890/09-2174.1.
- 67. Fox, R.J.; Bellwood, D.R. Niche partitioning of feeding microhabitats produces a unique function for herbivorous rabbitfishes (Perciformes, Siganidae) on coral reefs. *Coral Reefs* **2012**, *32*, 13–23, doi:10.1007/s00338-012-0945-5.
- Brandl, S.J.; Bellwood, D.R. Microtopographic refuges shape consumer-producer dynamics by mediating consumer functional diversity. *Oecologia* 2016, 182, 203–217, doi:10.1007/s00442-016-3643-0.
- 69. Fox, R.J.; Bellwood, D.R. Quantifying herbivory across a coral reef depth gradient. *Mar. Ecol. Prog. Ser.* 2007, 339, 49–59, doi:10.3354/meps339049.
- 70. Russ, G. Distribution and abundance of herbivorous grazing fishes in the central Great Barrier Reef. I. Levels of variability across the entire continental shelf. *Mar. Ecol. Prog. Ser.* **1984**, *20*, 23–34, doi:10.3354/meps020023.
- Wismer, S.; Hoey, A.; Bellwood, D.R. Cross-shelf benthic community structure on the Great Barrier Reef: Relationships between macroalgal cover and herbivore biomass. *Mar. Ecol. Prog. Ser.* 2009, 376, 45–54, doi:10.3354/meps07790.
- Choat, J.H.; Bellwood, D.R. Interactions amongst herbivorous fishes on a coral reef: Influence of spatial variation. *Mar. Biol.* 1985, *89*, 221–234, doi:10.1007/bf00393655.
- 73. Klumpp, D.; McKinnon, A. Temporal and spatial patterns in primary production of a coral-reef epilithic algal community. *J. Exp. Mar. Biol. Ecol.* **1989**, *131*, 1–22, doi:10.1016/0022-0981(89)90008-7.
- 74. Russ, G.R. Grazer biomass correlates more strongly with production than with biomass of algal turfs on a coral reef. *Coral Reefs* **2003**, *22*, 63–67, doi:10.1007/s00338-003-0286-5.
- 75. Mumby, P.J.; Harborne, A.R.; Williams, J.; Kappel, C.V.; Brumbaugh, D.R.; Micheli, F.; Holmes, K.E.; Dahlgren, C.P.; Paris, C.B.; Blackwell, P.G. Trophic cascade facilitates coral recruitment in a marine reserve. *Proc. Natl. Acad. Sci. USA* **2007**, 104, 8362–8367.
- 76. Bauman, A.G.; Seah, J.C.L.; Januchowski-Hartley, F.A.; Hoey, A.S.; Fong, J.; Todd, P.A. Fear effects associated with predator presence and habitat structure interact to alter herbivory on coral reefs. *Biol. Lett.* 2019, *15*, 20190409, doi:10.1098/rsbl.2019.0409.
- 77. Fong, C. R.; Sura, S. A.; Ford, A. T.; Howard, H. B.; Molina, N. E.; Smith, N. N.; Fong, P. Testing the conceptual and operational underpinnings of field herbivory assays: Does variation in predictability of resources, assay design, and deployment method affect outcomes?. J. Exp. Mar. Biol. Ecol. 2020, 533, 151469, doi:10.1016/j.jembe.2020.151469.
- Steinberg, P.; Paul, V. Fish feeding and chemical defenses of tropical brown algae in Western Australia. *Mar. Ecol. Prog. Ser.* 1989, 58, 253–259, doi:10.3354/meps058253.
- 79. Stiger, V.; Deslandes, E.; Payri, C.E. Phenolic contents of two brown algae, *Turbinaria ornata* and *Sargassum mangarevense* on Tahiti (French Polynesia): Interspecific, ontogenic and spatio-temporal variations. *Bot. Mar.* **2004**, *47*, doi:10.1515/bot.2004.058.
- 80. Bittick, S.J.; Bilotti, N.D.; Peterson, H.A.; Stewart, H.L. *Turbinaria ornata* as an herbivory refuge for associate algae. *Mar. Biol.* **2009**, 157, 317–323, doi:10.1007/s00227-009-1319-6.
- 81. Prathep, A.; Wichachucherd, B.; Thongroy, P. Spatial and temporal variation in density and thallus morphology of *Turbinaria ornata* in Thailand. *Aquat. Bot.* 2007, *86*, 132–138, doi:10.1016/j.aquabot.2006.09.011.
- 82. Seah, J.C.L.; Bauman, A.G.; Todd, P.A. Temporal variation in macroalgal removal: Insights from an impacted equatorial coral reef system. *Mar. Biol.* **2021**, *168*, 1–12, doi:10.1007/s00227-020-03806-7.
- 83. Clements, K.D.; German, D.P.; Piché, J.; Tribollet, A.; Choat, J.H. Integrating ecological roles and trophic diversification on coral reefs: Multiple lines of evidence identify parrotfishes as microphages. *Biol. J. Linn. Soc.* **2016**, doi:10.1111/bij.12914.
- 84. Perpetua, M.D.; Gorospe, J.G.; Torres, M.A.J.; Demayo, C.G. Diet composition based on stomach content of the Streaked spinefoot (*Siganus javus*) from three coastal bays in Mindanao, Philippines. *Adv. Environ. Sci.* **2013**, *5*, 49–61.
- 85. Miyake, S.; Ngugi, D.K.; Stingl, U. Diet strongly influences the gut microbiota of surgeonfishes. *Mol. Ecol.* **2015**, *24*, 656–672, doi:10.1111/mec.13050.

- Ebrahim, A.; Martin, T.S.H.; Mumby, P.J.; Olds, A.D.; Tibbetts, I.R. Differences in diet and foraging behaviour of commercially important rabbitfish species on coral reefs in the Indian Ocean. *Coral Reefs* 2020, *39*, 977–988, doi:10.1007/s00338-020-01918-6.
- 87. Nanami, A. Spatial distributions, feeding ecologies, and behavioral interactions of four rabbitfish species (*Siganus unimaculatus*, *S. virgatus*, *S. corallinus*, and *S. puellus*). *PeerJ* **2018**, *6*, e6145, doi:10.7717/peerj.6145.
- Bauman, A.G.; Hoey, A.S.; Dunshea, G.; Feary, D.A.; Low, J.; Todd, P.A. Macroalgal browsing on a heavily degraded, urbanized equatorial reef system. *Sci. Rep.* 2017, 7, 1–8, doi:10.1038/s41598-017-08873-3.
- Kuriiwa, K.; Hanzawa, N.; Yoshino, T.; Kimura, S.; Nishida, M. Phylogenetic relationships and natural hybridization in rabbitfishes (Teleostei: Siganidae) inferred from mitochondrial and nuclear DNA analyses. *Mol. Phylogenetics Evol.* 2007, 45, 69–80, doi:10.1016/j.ympev.2007.04.018.
- Fox, R.J.; Sunderland, T.L.; Hoey, A.; Bellwood, D.R. Estimating ecosystem function: Contrasting roles of closely related herbivorous rabbitfishes (Siganidae) on coral reefs. *Mar. Ecol. Prog. Ser.* 2009, 385, 261–269, doi:10.3354/meps08059.
- 91. Cheal, A.; Emslie, M.; Miller, I.; Sweatman, H. The distribution of herbivorous fishes on the Great Barrier Reef. *Mar. Biol.* 2012, 159, 1143–1154, doi:10.1007/s00227-012-1893-x.
- 92. Bellwood, D.R.; Choat, J.H. A functional analysis of grazing in parrotfishes (family Scaridae): The ecological implications. *Environ. Boil. Fishes* **1990**, *28*, 189–214, doi:10.1007/bf00751035.
- Brandl, S.J.; Rasher, D.B.; Côté, I.M.; Casey, J.M.; Darling, E.S.; Lefcheck, J.S.; Duffy, J.E. Coral reef ecosystem functioning: Eight core processes and the role of biodiversity. *Front. Ecol. Environ.* 2019, 17, 445–454, doi:10.1002/fee.2088.
- Fong, C.R.; Chancellor, K.S.; Renzi, J.J.; Robinson, D.R.; Barber, P.H.; Habtes, S.Y.; Fong, P. Epibionts on *Turbinaria ornata*, a secondary foundational macroalga on coral reefs, provide diverse trophic support to fishes. *Mar. Environ. Res.* 2018, 141, 39–43, doi:10.1016/j.marenvres.2018.08.001.
- 95. Clements, K. D.; Choat, J. H. Comparison of herbivory in the closely-related marine fish genera Girella and Kyphosus. *Mar. Biol.* **1997**, 127,4, 579-586.
- 96. Puk, L.; Cernohorsky, N.; Marshell, A.; Dwyer, J.; Wolfe, K.; Mumby, P. Species-specific effects of herbivorous fishes on the establishment of the macroalga Lobophora on coral reefs. *Mar. Ecol. Prog. Ser.* **2020**, 637, 1–14, doi:10.3354/meps13262.
- 97. Fulton, C.J.; Abesamis, R.A.; Berkström, C.; Depczynski, M.; Graham, N.A.; Holmes, T.H.; Tinkler, P. Form and function of tropical macroalgal reefs in the Anthropocene. *Funct. Ecol.* **2019**, *33*, 989–999.